# N<sub>x</sub>O emissions in response to the irrigation lower limits under different irrigation modes in a lettuce field

Maomao Hou<sup>1</sup>, Ying Xiao<sup>1</sup>, Qinyuan Zhu<sup>2</sup>, Jingnan Chen<sup>3</sup>, Huan Huang<sup>4</sup>, Qiu Jin<sup>5</sup>, Lin Zhu<sup>2\*</sup>, Fenglin Zhong<sup>1\*</sup>

(1. College of Horticulture, Fujian Agriculture and Forest University, Fuzhou 350000, China;

2. Nanjing Institute of Environmental Sciences, Nanjing 210000, China;

3. College of Horticulture and Gardening, Fujian Agricultural Vocational and Technical College, Fuzhou 350000, China;

4. Juhuang Group Co., Ltd., Xiamen 361021, China;

5. Nanjing Hydraulic Research Institute, Nanjing 210000, China)

**Abstract:** Irrigation has a significant impact on  $N_xO$  ( $N_2O$  and NO) emissions from cultivated land, yet the  $N_2O$  or NO emission among the irrigation lower limits under different irrigation modes has not been well compared. In an irrigated lettuce field, three DR (drip irrigation) lower limits were designed, including 75% (DR1), 65% (DR2) and 55% (DR3) field capacity, and one FI (furrow irrigation) lower limit (65% field capacity). The  $N_2O$  and NO emission fluxes and soil nitrogen (N) forms were determined, and the linear correlation between these indicators was analyzed. Results showed that under the same irrigation regime, the  $N_2O$  and NO emissions from furrow irrigation treatment. The cumulative  $N_2O$  and NO emissions under DR3 were 30.2% and 28.6% higher than under DR1, respectively. Moreover, DR1 was also the lowest among the four treatments in soil  $NO_3^-$ -N concentration. The  $N_2O$  and NO emission fluxes were more correlated to soil  $NH_4^+$ -N (r=0.88 and 0.76) or  $NO_2^-$ -N (r=0.90 and 0.80) concentration than soil  $NO_3^-$ -N and soluble organic N, indicating that  $N_2O$  and NO emission compared to that with drip irrigation regime was the same,  $N_2O$  and NO emissions were lower with drip irrigation treatment with small quota but high frequency reduced  $N_2O$  and NO emission compared to that with large quota but low frequency.

**Keywords:** nitrous oxide, nitric oxide, drip irrigation, furrow irrigation, soil nitrogen **DOI:** 10.25165/j.ijabe.20231604.7961

**Citation:** Hou M M, Xiao Y, Zhu Q Y, Chen J N, Huang H, Jin Q, et al. N<sub>x</sub>O emissions in response to the irrigation lower limits under different irrigation modes in a lettuce field. Int J Agric & Biol Eng, 2023; 16(4): 159–166.

#### 1 Introduction

Climate change is a hot social issue attracted international concern. A study predicts that the global average temperature will increase by 2°C in 2025 and 4°C in 2100 compared with 100 years ago<sup>[1]</sup>. N<sub>2</sub>O is one of the main greenhouse gases, and its global warming potential is about 300 times than that of CO<sub>2</sub>. NO participates in photochemical reactions, destroys the ozone layer, and indirectly causes the greenhouse effect<sup>[2]</sup>. Agricultural production process emits a great amount of N<sub>2</sub>O and NO gas<sup>[3]</sup>. Taking N<sub>2</sub>O as example, the amount of N<sub>2</sub>O emitted by global cultivated land is about  $6.4 \times 10^{12}$  g N, accounting for one fourth of the global total N<sub>2</sub>O emissions<sup>[4]</sup>. Agricultural management measures, such as irrigation, fertilization, mulching and ploughing,

are proved to have an impact on N2O and NO emissions<sup>[5-8]</sup>.

Irrigation is an indispensable agricultural management measure. Many studies showed that the irrigation amount (irrigation quota) affects the emission of nitrogen oxides from agricultural land<sup>[9]</sup>. The research by Chen et al.<sup>[10]</sup> showed that the soil N<sub>2</sub>O emission increases gradually as irrigation quota increased, N<sub>2</sub>O emission under 60% and 80% irrigation quotas decreased by 19.1% and 8.0% respectively compared to under full irrigation quota. Du et al.<sup>[11]</sup> employed the static box in-situ gas sampling method and found that average cumulative N<sub>2</sub>O emission by full irrigation is 1.27 times than that by deficit irrigation.

In addition, irrigation modes (drip irrigation, flood irrigation, sprinkler irrigation, infiltration irrigation, etc.) also influence the  $N_2O$  and NO emissions<sup>[12]</sup>. The irrigation modes change the soil pore structure, affect the diffusion of  $O_2$  in the soil, and then affect the soil microbial activities<sup>[13]</sup>. The nitrification and denitrification processes dominated by soil microorganisms are influenced, this finally impacting  $N_2O$  and NO emissions<sup>[14]</sup>. Moreover, different irrigation modes lead to different water and N distributions in the surface and profile soil, and have different effects on the leaching and mineralization of soil N<sup>[15-17]</sup>, which will inevitably change the amount and location of available N in surface soil. Available N in surface soil is an important substrate responsible for N<sub>2</sub>O and NO emissions<sup>[18]</sup>.

In general, previous studies regarding the impact of irrigation on  $N_2O$  and NO emissions mainly focused on the irrigation quotas or irrigation modes, little attention has been paid to the irrigation

Received date: 2022-10-10 Accepted date: 2023-06-15

Biographies: Maomao Hou, PhD, Associate Professor, research interest: agricultural water conservancy engineering, Email: xiaohouboshi@126.com; Ying Xiao, MS candidate, research interest: agricultural water management, Email: xiaoying0112@126.com; Qinyuan Zhu, MS candidate, research interest: environmental engineering, Email: zhuqinyuan569@163.com; Jingnan Chen, PhD, research interest: agricultural water conservancy engineering, Email: 53cjn@sina.com; Huan Huang, MS candidate, research interest: environmental engineering, Email: 2441694627@qq.com; Qiu Jin, PhD, research interest: watersaving irrigation and soil improvement, Email: fengyuwuzujq@163.com.

<sup>\*</sup>Corresponding author: Lin Zhu, MS candidate, research interest: environmental engineering, Email: zhulin@nies.org; Fenglin Zhong, Professor, research interest: crop cultivation. College of Horticulture, Fujian Agriculture and Forest University, Tel: +86-15659385002, Email: faczhong@163.com.

frequency (controlled by different irrigation lower limits). The difference of  $N_2O$  or NO emission between drip and furrow irrigation under the same irrigation frequency was also rarely studied. This study designed three different lower limits of DR (drip irrigation) and one lower limit of FI (furrow irrigation), to investigate the regularity of  $N_2O$  and NO emissions under different treatments, and to analyze the relationship between  $N_2O$  (NO) emission and soil N forms. The objectives were: (1) to compare the  $N_2O/NO$  emission between DR and FI under the same irrigation lower limit; (2) to compare the  $N_2O/NO$  emission among different drip irrigation lower limits; (3) to explore the driving mechanism of  $N_2O$  and NO emission by the irrigations.

#### 2 Materials and methods

#### 2.1 Experimental site

The experiment was conducted at Fruit Science and Technology Demonstration Base from September 20 of 2021 to January 6 of 2022 in Yunxiao County, Fujian Province of China. Yunxiao County is under the subtropical marine monsoon climate. The frost free period is 347 d. The extreme maximum and the extreme minimum temperature are 38.1°C and -0.2°C, respectively. The annual average temperature is 21.3°C and the annual accumulated temperature is 7548.8°C. The annual precipitation in the experimental area is 1730.6 mm. The soil variety is the ferrallitic soil. The physical and chemical properties of soils in plough layer are as follows: pH of 5.9, bulk density of 1260 kg/m<sup>3</sup>, field water capacity of 29.8%, organic matter of 3.45%, available N of 90.2 mg/kg, available P of 12.2 mg/kg and available K of 152.3 mg/kg.

#### 2.2 Experimental design

The lettuce variety Feigiao lettuce No. 1 was used as the experimental material. The substrate was used to cultivate seedlings. When these seedlings grew out 4-5 expanded leaves (October 15), they were transplanted into the soils. The rosette stage and fleshy stem expansion stage are the two stages when lettuce needs major amounts of water, and these two periods are the water sensitive periods for lettuce. Therefore, after transplanting, three different DR lower limits (75% (DR1), 65% (DR2) and 55% (DR3) field capacity) and one FI lower limit (65% field capacity) were arranged from the rosette stage for the lettuces. The soil moisture content of the cultivated layer was measured every day used for finding the lower limits. Once the moisture content reached the lower limit, the irrigation was started. The upper limit of all the treatments was 95%. For each time, the irrigation amount of FI was the same as that of DR2. The local irrigation habit was the furrow irrigation, and the irrigation water was applied to 1/3 of the furrow height then dried naturally. In this study, the converted furrow irrigation quota according to DR2 was consistent with the local practice. The irrigation quota was calculated as<sup>[19]</sup>:

$$M = S \times r \times h \times Q \times (q^1 - q^2)/0.95 \tag{1}$$

where, *M* is the irrigation quota,  $m^3$ ; *S* is the irrigation area,  $m^2$ ; *r* is the soil bulk density, kg/m<sup>3</sup>; *h* is the planned depth of wetted soil (0.2 m); *q* is field water capacity, %; *q*<sup>1</sup> is the upper irrigation limit (95%); *q*<sup>2</sup> is the lower irrigation limit (75%, 65% or 55%), and 0.95 is the irrigation coefficient.

To sum up, there were four different treatments in this study, and each treatment was repeated three times. The division of lettuce growth stages and the treatment periods are listed in Table 1. The irrigation parameters are listed in Table 2.

Table 1	Growth	stages	of lettuce
т яппет	1 1 1 1 1 1 1 1 1	NIXUES.	or rennee

Date	Growth stages	Irrigation treated periods
2021.9.20-2021.9.22	Seed germination	
2021.9.23-2021.10.15	Seedling stage	
2021.10.16-2021.11.21	Rosette stage	${\sim}$
2021.11.22-2022.1.2	Fleshy stem expansion stage	$\stackrel{\sim}{\sim}$
2022.1.3-2022.1.6	Harvest stage	

Note:  $\Rightarrow$  represents that in this period, the lettuce was treated with different irrigation treatments (three drip irrigation lower limits (75%, 65%, 55% field capacity) and one furrow irrigation lower limit (65% field capacity).

Table 2	The	irrigation	regime
I abic L	1 110	maganon	1 CEIIIIC

	Rosette stage to fleshy stem expansion stage				The irrigation amount
Treatment	Irrigation times	Irrigation interval/d	Irrigation quota/mm	Irrigation amount/mm	during whole growth period/mm
DR1	8	9.5	15.8	126.5	284.1
DR2	5	15.2	23.7	118.6	276.2
DR3	3	25.3	31.6	94.9	252.5
FI	5	15.2	23.7	118.6	276.2

Field experimental area was separated into 12 blocks. Each treatment took up 3 blocks, and each block occupied an area of 8 m×4 m. In one block, three ridges of lettuces were cultivated, with the ridge height of 20 cm and width of 60 cm. The spacing between two ridges was 20 cm (Figure 1). The lettuce row-to-row spacing was 30 cm and plant-to-plant spacing was 35 cm. An additional soil ridge without lettuce cultivation was arranged between the furrow irrigation block and other block.





The fertilization management for each block was the same. Total fertilizer application rates were 675 kg/hm<sup>2</sup> of urea, 600 kg/hm<sup>2</sup> of calcium superphosphate and 375 kg/hm<sup>2</sup> of potassium sulfate. All the calcium superphosphate was applied as basal fertilizer, whereas the urea, as well as the potassium sulfate, was applied according to 40% for basal fertilizer, 20% for first topdressing and 40% for second topdressing. The dates for basal fertilization, first topdressing and second topdressing were October 15, November 1 and November 23, respectively. The fertilizers were applied into the 6 cm soil depth using a hole applicator. In early and late growth stages of lettuce, the heart rot and the downy mildew were prevented respectively using plant protection chemicals including propineb, imidacloprid, putrescine, etc.

#### 2.3 Sampling and measurement

The gas collection device is a self-made cylindrical static cover (composed of cover and base, Figure 2). The cover body is made of PVC and enclosed with a layer of silver reflective film. The diameter of the cover body is 30cm and the height is 80 cm. The base can be buried in the soil, and the upper part is provided with a water tank to play a sealing role. Before collecting, the water tank is supplied with clean water, and then the cover is placed on the base. A thermometer and a gas collection port are installed on the side of the cover. A small fan driven by the external battery is installed at the top of the cover to make the emitted gas distributed uniformly.



Figure 2 The gas sampling device

One fixed gas sampling point is set for each block. Gas samples were collected every 7 d from the second day after transplanting. For the whole growth stage of lettuce, 13 times of the gas sampling were conducted. Each collection started when there was no obvious water on the soil surface. The collection time was from 8:00 to 9:00 in the morning. The gas was collected at 0, 10, 20 and 30 min after sealing and the temperature inside the cover was recorded. A 50 mL syringe was used to extract gas and then injected into a 40 mL gas collecting bottle. The gas collecting bottles were taken to the laboratory to measure N<sub>2</sub>O emission flux using a gas chromatograph (Agilent 7890B, Agilent Technologies, USA). At the same time, 1000 mL gas was collected into an airtight bag to measure the NO emission flux with a 42i NO-NO<sub>2</sub>-NO<sub>X</sub> analyzer (Thermo Environmental Instruments Inc., USA).

For each gas collection, soil samples in 0-10 cm soil layer were collected synchronously. The soil collection position was at the bottom of the base. After multi-point collection, the soils were mixed evenly for measuring the soil mass moisture content,  $NO_3^-$ -N,  $NH_4^+$ -N,  $NO_2^-$ -N and soluble organic N. The soil moisture content was measured by drying method<sup>[20]</sup>. The soluble total N was determined by the alkaline potassium persulfate oxidation method<sup>[21]</sup>. The  $NO_3^-$ -N,  $NH_4^+$ -N and  $NO_2^-$ -N were extracted using 0.01 mol/L CaCl<sub>2</sub><sup>[22]</sup> then determined by an automatic analyzer (Seal Analytical, USA).

#### 2.4 Data analysis

The N<sub>2</sub>O and NO emission fluxes are calculated as follows<sup>[23]</sup>:

$$F = \rho \cdot H \cdot \frac{273}{273 + T} \cdot \frac{\mathrm{d}c}{\mathrm{d}t} \tag{2}$$

where, *F* is N<sub>2</sub>O or NO emission flux,  $\mu$ g N/m<sup>2</sup>·h;  $\rho$  is the density of N<sub>2</sub>O and NO in the standard state, 1.25 and 0.625 kg/m<sup>3</sup>; *H* is the effective height of the cover, m; *T* is the actual temperature inside the cover at the moment of measurement, °C; d*c*/d*t* is the change rate of N<sub>2</sub>O or NO concentration, mL/L·h. Since the air pressure inside the cover is nearly constant, the impact of the air pressure on N<sub>2</sub>O or no is ignored.

The cumulative emission of  $N_2O$  or NO is calculated according to the following equations<sup>[24,25]</sup>:

$$M = \sum (F_{i+1} + F_i)/2 \times (t_{i+1} - t_i) \times 24 \times 10^{-5}$$
(3)

where, *M* is the cumulative emission of N<sub>2</sub>O or NO, kg N/hm<sup>2</sup>; *F* is the  $i^{\text{th}}$  N<sub>2</sub>O or NO emission flux,  $\mu$ g N/m<sup>2</sup>·h; *i* is the number of measurements;  $t_{i+1}-t_i$  is the days between two determinations, d.

Soil soluble organic N is calculated according to [26]:

$$SON = STN - MN \tag{4}$$

where, SON is soil soluble organic N, mg/kg, STN is soluble total N,

mg/kg, MN is mineral N, including NO<sub>3</sub><sup>-</sup>-N, NH<sub>4</sub><sup>+</sup>-N and NO<sub>2</sub><sup>-</sup>-N.

The soil water filled pore spaces (*WFPS*, %) is calculated according to [27]:

$$SWPS = \left(\frac{\rho}{1-\rho/2.65}\right) \cdot Q_{\rm w} \tag{5}$$

where,  $\rho$  is the soil bulk density, g/cm<sup>3</sup>; 2.65 is the soil density, g/cm<sup>3</sup>;  $Q_w$  is the soil mass moisture content, %.

#### 2.5 Statistical analysis

Significant differences among different treatments were calculated according to Duncan's multiple range test using SPSS 17.0 software.

#### **3** Results and analysis

### 3.1 Dynamic changes of WFPS under different irrigation treatments

During the whole growth period of lettuce, the WFPS showed the fluctuating variation trend, and there were obvious differences of WFPS under different irrigation treatments (Figure 3). The WFPS under DR1 was in a high level, reaching 54.5%-62.9%; followed by DR2, reaching 49.2%-59.6%; DR3 was relatively lower, recording as 41.1%-63.4%. The variation trend of WFPS under DR2 and FI was similar, whereas the WFPS under DR2 was slightly higher than that under FI during the whole lettuce growth period. The maximum WFPS difference occurred on the 71 d after transplanting, and was 47.0% between DR1 and DR3. The average WFPS under DR1, DR2, DR3 and FI for the whole growth period was 58.5%, 53.9%, 47.8% and 52.2% respectively, and the maximum difference of average WFPS reached 22.4%.



Note: DR1, DR2 and DR3 represent lower irrigation limits of 75%, 65% and 55% accounted for the field capacity, respectively. FI is the furrow irrigation treatment with the same irrigation regime as DR2.

Figure 3 Variation of WFPS under different irrigation treatments

#### 3.2 Impact of different irrigation treatments on N<sub>2</sub>O emission

The maximum N<sub>2</sub>O emission flux appeared at the second day after transplanting, reaching 358.5-1000.2  $\mu$ g N/m<sup>2</sup>·h (Figure 4a). From 2 to 15 d after transplanting, the N<sub>2</sub>O emission flux decreased significantly and then gradually stabilized. At 22 and 43 d after transplanting, the N<sub>2</sub>O emission flux appeared two small peaks. The 22 d and 43 d after transplanting exactly corresponded to the second and third fertilization dates. In general, the N<sub>2</sub>O emission flux of FI in the whole growth period was at the highest level among different treatments, while that of DR1 was overall the lowest.

There were two growth stages for the increase of cumulative  $N_2O$  emission (Figure 4b): the rapid growth stage was from 2 to 22 d after transplanting, and the relatively slow growth stage was occurred from 22 to 83 d after transplanting. The difference of the



Note: DR1, DR2 and DR3 represent lower irrigation limits of 75%, 65% and 55% accounted for the field capacity, respectively. FI is the furrow irrigation treatment with the same irrigation regime as DR2. The Data were mean±SD.

Figure 4 Effect of different irrigation treatments on N<sub>2</sub>O emission flux and cumulative N<sub>2</sub>O emission

accumulated N<sub>2</sub>O emission among treatments began to be obvious from 8 d after transplanting. At 8 d after transplanting, the accumulated N<sub>2</sub>O emission under FI was the highest, which was nearly three times than that under DR1. At 83 d after transplanting, the last measurement, it was found that the cumulative N<sub>2</sub>O emission under DR1, DR2, DR3 and FI were 1.50, 1.77, 2.15 and 2.80 kg N/hm<sup>2</sup>. This suggested that under the same irrigation regime, the N<sub>2</sub>O emission under furrow irrigation was 36.8% higher compared to that under drip irrigation. In addition, DR3 increased N<sub>2</sub>O emissions by 30.2% compared with DR1, indicating that different irrigation regimes but same irrigation mode obviously affect the cumulative N<sub>2</sub>O emission.

#### 3.3 Impact of different irrigation treatments on NO emission

The variation regularity of NO emission flux was similar to that of N<sub>2</sub>O (Figure 5a). After decreasing in early growth stage of lettuce, it tended to be stable from 29 d after transplanting. The difference of NO emission flux among the treatments was mainly detected in the early growth stage, especially during 2-29 d after transplanted. At harvest stage, the NO emission flux decreased to about one tenth of the highest. Among the different treatments, the NO emission flux under DR1 during the whole growth period was at a lower level, and the highest NO emission flux under DR1 occurred on the second day after transplanting, which was 101.4  $\mu$ g N/m<sup>2</sup>·h, less than half under FI at the same measurement time.



Note: DR1, DR2 and DR3 represent lower irrigation limits of 75%, 65% and 55% accounted for the field capacity, respectively. FI is the furrow irrigation treatment with the same irrigation regime as DR2.

Figure 5 Effect of different irrigation treatments on NO emission flux and cumulative NO emission

Cumulative NO emissions under DR1 and DR2 were not significantly different (Figure 5b), and both were obviously lower compared to under DR3 or FI. At 83 d after transplanting, the cumulative NO emissions under DR1, DR2, DR3 and FI were 0.60, 0.63, 0.84 and 1.15 kg N/hm<sup>2</sup>. The average NO emission under drip irrigation was 0.69 kg N/hm<sup>2</sup>, which was 40% less than that under FI. With the same irrigation regime, DR2 emitted 45.2% less NO in comparison to FI. Among the different drip irrigation treatments, DR3 emitted 28.6% more NO compared to DR1.

#### 3.4 Impact of different irrigation treatments on soil N

In general, the contents of different forms of soil N from high

to low were soil soluble organic N,  $NO_3^-$ -N,  $NH_4^+$ -N and  $NO_2^-$ -N (Figures 6a-6d). The overall variation regularity was that three peaks were found at 2 d, 22 d and 43 d after transplanting. This regularity was more obvious on soil  $NH_4^+$ -N and  $NO_2^-$ -N. There were differences in soil  $NO_3^-$ -N among the treatments, but this difference was not regular at each measurement, except that the  $NO_3^-$ -N content under DR1 was generally in the lowest level among the four treatments. The soil  $NH_4^+$ -N or  $NO_2^-$ -N content under FI and DR3 was higher than under DR1 and DR2 treatments. The average soil  $NH_4^+$ -N contents under FI, DR3, DR2 and DR1 during the whole growth period was 55.3, 51.1, 48.9 and 43.0 mg/kg, and





Figure 6 Effect of different irrigation treatments on soil  $NO_3^-$ -N,  $NH_4^+$ -N,  $NO_2^-$ -N and soluble organic N in 0-10 cm soil layer

the average  $NO_2^-$ -N contents was 0.97, 0.80, 0.68 and 0.59 mg/kg. Compared with the other three different forms of N, soil soluble organic N was at a more stable level, and the variation range was relatively small. DR1 and DR2 have almost no difference in soil soluble organic, but both were lower than DR3.

# 3.5 Correlation analysis between $N_2O$ (NO) emission and possible impact factors

The N<sub>2</sub>O emission flux has a positive correlation with soil  $NH_4^+$ - N, with a correlation coefficient of 0.88 (Figure 7). Meanwhile, a stronger positive correlation between N<sub>2</sub>O emission flux and NO<sub>2</sub><sup>-</sup>-



Figure 7 Correlation between N<sub>2</sub>O emission and possible impact factors

N was observed, and the correlation coefficient reached 0.90. However, the correlation between  $N_2O$  emission flux and  $NO_3^-$ -N or soluble organic N was not obvious.

Similarly, the NO emission fluxes were positively correlated with  $NH_4^+$ -N or  $NO_7^-$ -N, with correlation coefficients of 0.76 and

0.80 respectively (Figure 8). However, the correlation coefficient between NO emission flux and  $NO_3^-$ -N or soluble organic N was only 0.04 and 0.17, indicating that NO emission flux has no obvious relationship with  $NO_3^-$ -N or soluble organic N in the surface soil.



Figure 8 Correlation between NO emission and possible impact factors including soil NO<sub>3</sub><sup>-</sup>-N, NH<sub>4</sub><sup>+</sup>-N, NO<sub>2</sub><sup>-</sup>-N and soluble organic N

#### 4 Discussion

## 4.1 $N_2O$ and NO emission in response to different irrigation schemes

Soil moisture regulates the nutrient content in soil pore and affects the nitrification and denitrification processes<sup>[28,29]</sup>. In this study, the N<sub>2</sub>O and NO emissions under DR1 treatment were lower than that under DR3, which was in line with the results by Zhang<sup>[18]</sup>. This might be due to the fact that the soils in DR3 experienced a more intense dry-wet alternation process, leading to: (1) the promotion in the mineralization of soil N and the formation of more substrates  $(NH_4^+-N \text{ or } NO_3^--N)^{[30,31]}$ ; (2) the more intension in the alternating nitrification and denitrification process<sup>[32]</sup>. Both the two results will increase the emission of N2O and NO. In addition, the reason responsible for the higher NO emission under DR3 might be that NO was mainly caused by nitrification process driven by autotrophic and heterotrophic nitrifying bacteria<sup>[33]</sup>, the soil WFPS under DR3 was in a lower level among the three drip treatments (Figure 3), this was conducive to the occurrence of the nitrification process. Abalos et al.<sup>[34]</sup> also found that compared with highfrequency irrigation, the low-frequency irrigation increased NO emission

In this study, it was found that under the same irrigation

regime, drip irrigation decreased the N2O and NO emissions compared to furrow irrigation, which agreed with most previous studies<sup>[35-37]</sup>. Andrew et al.<sup>[38]</sup> found that the drip irrigation reduced soil N<sub>2</sub>O by up to 62%. Wang et al.<sup>[39]</sup> used three irrigation modes including surface drip irrigation, sprinkler irrigation and furrow irrigation for the winter wheat in north China plain, and found that the N<sub>2</sub>O emission from drip irrigation treatment was 14.6% lower than that from furrow irrigation. Laura<sup>[35]</sup> compared the difference of N<sub>2</sub>O emission from soil under furrow irrigation and drip irrigation during melon cultivation season, and showed that the drip irrigation reduced 70% of total N2O emissions compared with furrow irrigation. The reason might be that the water rapidly filled the soil voids in a short time under furrow irrigation, forming an anaerobic environment, strengthening denitrification process and promoting N<sub>2</sub>O emission; meanwhile, the soil treated by furrow irrigation obtained greater wetted volume and more microorganisms with stronger activity<sup>[40]</sup>, which might also be another reason for promoting denitrification thus increasing the N2O emission. After furrow irrigation, soil water was easier to downward infiltrate compared to drip irrigation, and the soil in the cultivated layer dehydrated faster<sup>[41]</sup>, conversely forming a nitrification environment conducive to NO emission. On the contrary, some researchers achieved different conclusions. For example, Guo et al.[42] found that the average N<sub>2</sub>O emission flux reached 74.81  $\mu$ g/m<sup>2</sup>·h under drip irrigation in wheat field, and was increased by 25.87% compared with furrow irrigation. According to the analysis by earlier researchers, there might have two main reasons: firstly, drip irrigation reduced the leaching of N, and the total amount of mineral N in surface soil was greater, and correspondingly, the substrate concentration for N<sub>2</sub>O and NO emissions was higher<sup>[43]</sup>; secondly, under drip irrigation, the soil structure was more intact and the available oxygen content was higher, which strengthened the mineralization of organic N<sup>[44]</sup> and then promoted the nitrification and denitrification processes, finally increased the production and emission of N<sub>2</sub>O and NO from the soil.

#### 4.2 Impact factors responsible for N<sub>2</sub>O and NO emission

In this study, it was found that different lower irrigation limits have different degrees of impact on soil N ( $NO_3^--N$ ,  $NO_4^+-N$ ,  $NO_2^--N$ . etc), and DR3 treatment was at the highest level of mineral N in the whole growth period among the three drip irrigation treatments. There might be two reasons for this: (1) DR3 has the lowest irrigation volume and frequency, which reduced the downward leaching amount of N from the arable soil<sup>[45]</sup>; (2) the soil under DR3 was in a drier state (Figure 3), the diffusion of soil nutrients and ions was limited, less energy could be provided to microorganisms, thus the microorganisms were in a "hungry" state. After rehydration, microorganisms might grow and reproduce compensatorily, and their number and activity were significantly increased, thus promoting the mineralization of soil N<sup>[46]</sup>. The higher soil mineral N detected under DR3 (Figure 6) could support the above presumption.

The driving pathways for the soil  $N_2O$  and NO emissions included the above-mentioned nitrification and denitrification, as well as the denitrification by the nitrifying bacteria and the dissimilatory reduction of nitrate nitrogen to ammonium<sup>[47,48]</sup>. However, it was generally believed that nitrification and denitrification were the two main pathways<sup>[49]</sup>. Soil NO<sub>2</sub><sup>-</sup>-N was the product of ammonia oxidation during the nitrification process<sup>[50]</sup>. A significant positive correlation was observed between N<sub>2</sub>O/NO emission and soil NO<sub>2</sub><sup>-</sup>-N or NH<sub>4</sub><sup>+</sup>-N, indicating that NO<sub>2</sub><sup>-</sup>-N or NH<sub>4</sub><sup>+</sup>-N were the key factors to explain the changes of N<sub>2</sub>O and NO emission. Therefore, it could be speculated that N<sub>2</sub>O and NO emissions were dominated by the nitrification. Moreover, from the perspective of soil WFPS (Figure 3), the lower WFPS corresponded to the higher N<sub>2</sub>O or NO emission, which also supports that the driven pathway of N<sub>2</sub>O/NO emission was nitrification.

The study quantitatively compared the emission differences of N<sub>2</sub>O/NO under different drip irrigation lower limits, as well as under different irrigation modes. Furthermore, the correlation between N<sub>2</sub>O/NO emission and possible influencing factors was also analyzed. However, different irrigations may cause different distribution of water and N in soil profile<sup>(19)</sup>. Soil capillary action will promote water and nitrogen to accumulate to the surface soil, therefore affecting the NO<sub>x</sub> emission. Whether and to what extent, the movement of soil water and N in profile soil affects the emission of nitrogen oxides, need to be studied. In addition, irrigation influences the development of crop roots then impacts the plant absorption for soil N<sup>[51]</sup>. Does plant N consumption have an impact on N<sub>2</sub>O/NO emission? This is worthy to be investigated in future.

#### 5 Conclusions

The overall results concluded that under the same irrigation regime,  $N_2O$  and NO emissions from furrow irrigation treatment

increased by 36.8% and 45.2% respectively compared to that from drip irrigation treatment. Among the drip irrigation treatments, the cumulative N2O or NO emission was the lowest in DR1 and the highest in DR3. The cumulative N2O and NO emissions under DR3 were 30.2% and 28.6% higher than under DR1, respectively. Moreover, DR1 was also the lowest among the four treatments in soil NO<sub>3</sub><sup>-</sup>-N concentration. Both N<sub>2</sub>O and NO emission fluxes were positively correlated to soil NO<sub>4</sub><sup>+</sup>-N (r=0.88 and 0.76) or NO<sub>2</sub><sup>-</sup>-N (r=0.90 and 0.80) concentration, while not obviously correlated to NO3-N and soluble organic N, indicating that N2O and NO were mainly produced by the soil nitrification process. It was concluded that the N2O and NO emissions under drip irrigation were lower than under furrow irrigation, when the irrigation regime was the same. Besides, drip irrigation with "small quota but high frequency" reduced N<sub>2</sub>O and NO emission compared to that with "large quota but low frequency".

#### Acknowledgements

This work was financially supported by the cross fusion project "Green Civil Engineering and Water Conservancy" (000-71202103D); Young and Middle-aged Project of Education Department of Fujian Province (JAT210784); Natural Science Foundation of Fujian province (2023J01133579).

#### [References]

- [1] Xu B Y, Wang T J, Ma D Y, Song R, Zhang M, Gao L B, et al. Impacts of regional emission reduction and global climate change on air quality and temperature to attain carbon neutrality in China. Atmospheric Research, 2022; 279. doi:10.1016/J.ATMOSRES.2022.106384
- [2] Anenberg S, Schwartz ., Shindell D, Amann M, Faluvegi G, Klimont Z, et al. Global air quality and health co-benefits of mitigating near-term climate change through methane and black carbon emission controls. Environmental Health Perspectives, 2012; 120: 831–839.
- [3] Kleinen T, Gromov S, Steil B, Brovkin V. Atmospheric methane underestimated in future climate projections. Environmental Research Letters, 2022; 16: 122–131.
- [4] Marquina S, Pérez T, Donoso L, Giuliante A, Rasse R, Herrera F. NO, N<sub>2</sub>O and CO<sub>2</sub> soil emissions from Venezuelan corn fields under tillage and notillage agriculture. Nutrient Cycling in Agroecosystems, 2015; 101: 123–137.
- [5] Stehfest E, Bouwman L. N<sub>2</sub>O and NO emission from agricultural fields and soils under natural vegetation: summarizing available measurement data and modeling of global annual emissions. Nutrient Cycling In Agroecosystems, 2006; 74: 207–228.
- [6] Franco-Luesma S, Cavero J, Plaza D, Cantero C, Tortosa G, Bedmar E, et al. Irrigation and tillage effects on soil nitrous oxide emissions in maize monoculture. Agronomy Journal, 2020; 112: 56–71.
- [7] Lee S-I, Kim G-Y, Lee J S, Choi E J. Effect of no-tillage and green manure practices on the nitrous oxide emission from cropland. Korean Journal of Environmental Biology, 2019; 37: 309–316.
- [8] Osterholz W R, Kucharik C J, Hedtcke J L, Posner J L. Seasonal nitrous oxide and methane fluxes from grain- and forage-based production systems in Wisconsin, USA. Journal of Environmental Quality, 2014; 43: 1833–1843.
- [9] Tian D, Zhang Y Y, Mu Y J, Zhou Y Z, Zhang C L, Liu J F. The effect of drip irrigation and drip fertigation on N<sub>2</sub>O and NO emissions, water saving and grain yields in a maize field in the North China Plain. Science of the Total Environment, 2017; 575: 1034–1040.
- [10] Chen H, Shang Z H, Wang Y F, Zhu Y, Cai H J, Zhang Y L, et al. Effects of irrigation amounts on soil CO<sub>2</sub>, N<sub>2</sub>O and CH<sub>4</sub> emissions in greenhouse tomato field. Chinese Journal of Applied Ecology, 2019; 30: 102–111.
- [11] Du Y D, Zhang Q, Cui B J, Gu X B, Niu W Q. Effects of water and nitrogen coupling on soil N<sub>2</sub>O emission characteristics of greenhouse celery field under aerated irrigation. Transactions of the CSAE, 2017; 33: 127–134. (in Chinese)
- [12] Zhang Y J, Niu H S, Wang S P, Xu K, Wang R. Application of the DNDC model to estimate N<sub>2</sub>O emissions under different types of irrigation in vineyards in Ningxia, China. Agricultural Water Management, 2016; 163:

295-304.

- [13] Yu Y Z, Jiao Y, Yang W Z, Song C N, Zhang J, Liu Y B. Mechanisms underlying nitrous oxide emissions and nitrogen leaching from potato fields under drip irrigation and furrow irrigation. Agricultural Water Management, 2022; 260: 107270.
- [14] Ye X H, Han B, Li W, Zhang X C, Zhang Y L, Lin X G, et al. Effects of different irrigation methods on nitrous oxide emissions and ammonia oxidizers microorganisms in greenhouse tomato fields. Agricultural Water Management, 2018; 203: 115–123.
- [15] Hou M M, Jin Q, Lu X Y, Li J Y, Zhong H Z, Gao Y. Growth, Water use, and nitrate-n-15 uptake of greenhouse tomato as influenced by different irrigation patterns, N-15 labeled depths, and transplant times. Frontiers in Plant Science, 2017; 8: 666.
- [16] Hou M M, Zhong F L, Jin Q, Liu E J, Feng J, Wang T Y, et al. Fate of nitrogen-15 in the subsequent growing season of greenhouse tomato plants (Lycopersicon esculentum Mill) as influenced by alternate partial root-zone irrigation. Rsc Advances, 2017; 7: 34392–34400.
- [17] Wu Z J, Shi Y L, Li D P, Lu Z Y, Wei Z P, Zhang L L, Gong P, et al. Fate of urea-N-15 as influenced by different irrigation modes. RSC Advances, 2020; 10: 11317–11324.
- [18] Zhang L. Effects of Irrigation Regimes on N<sub>2</sub>O and NO Emissions from greenhouse soil. Scientia Agricultura Sinica, 2021; 54: 992–1002.
- [19] Chen J N, Jin Q, Zhu Q Y, Xiao Y, Zhu J W. Fate of soil residual fertilizern-15 as affected by different drip irrigation regimes. Water, 2022; 14: 365–375.
- [20] Jin Q, You J J, Xie M X, Qiu Y L, Lei S H, Ding Q, et al. Drip irrigation reduced fertilizer nitrogen loss from lettuce field-a case study based on N-15 tracing technique. Water, 2022; 14: 256–264.
- [21] Zhou L. Alkaline potassium persulfate oxidation method for measuring soil N. Physical and Chemical Analysis, 2017; 53: 137–140.
- [22] Shao X H, Hu A T, Houba V, Novozamsky I. 0.01 CaCl<sub>2</sub> as extration for different forms of soil nitrogen. Acta Pedologica Sinica, 1991; 12: 112–118.
- [23] Shao Y L, Chen J N, Wang L K, Hou M M, Chen D L. Effects of fermented organic fertilizer application on soil N<sub>2</sub>O emission under the vegetable rotation in polyhouse. Environmental Research, 2021; 200: 111491.
- [24] Ni B, Xu X C, Zhang W, Yang X, Liu R, Wang L G, et al. Reduced fertilization mitigates N<sub>2</sub>O emission and drip irrigation has no impact on N<sub>2</sub>O and NO emissions in plastic-shed vegetable production in northern China. Science of the Total Environment, 2022; 824: 153976.
- [25] Lan T, Zhang H, Han Y, Deng O P, Tang X Y, Luo L, et al. Regulating CH<sub>4</sub>, N<sub>2</sub>O, and NO emissions from an alkaline paddy field under rice–wheat rotation with controlled release N fertilizer. Environmental Science and Pollution Research, 2021; 28(14). doi:10.1007/S11356-020-11846-1.
- [26] Liang B, Yang X Y, He X H, Zhou J B. Effects of 17-year fertilization on soil microbial biomass C and N and soluble organic C and N in loessial soil during maize growth. Biology and Fertility of Soils, 2011; 47: 121–128.
- [27] Dobbie K E, Smith K A. The effect of water table depth on emissions of N<sub>2</sub>O from a grassland soil. Soil Use and Management, 2006; 22: 22–28.
- [28] Liu Y, Cong R H, Liao S P, Guo Q, Li X K, Ren T, et al. Rapid soil rewetting promotes limited N<sub>2</sub>O emissions and suppresses NH<sub>3</sub> volatilization under urea addition. Environmental Research, 2022; 212: 113402.
- [29] Rahmawati A, Neve S D, Purwanto B H. N<sub>2</sub>O-N Emissions from Organic and Conventional Paddy Fields from Central Java, Indonesia. Procedia Environmental Sciences, 2015; 28: 606–612.
- [30] Bagheri S, Mirseyed H, Etesami H, Razavipour T. Rice straw and composted azolla alter carbon and nitrogen mineralization and microbial activity of a paddy soil under drying-rewetting cycles. Applied Soil Ecology, 2020; 154: 103638.
- [31] Gao J Q, Feng J, Zhang X W, Yu F H, Xu X, Kuzyakovet Y, et al. Dryingrewetting cycles alter carbon and nitrogen mineralization in litter-amended alpine wetland soil. Catena, 2016; 145: 285–290.
- [32] Verhoeven E, Decock C, Barthel M, Bertora C, Sacco D, Romani M, et al. Nitrification and coupled nitrification-denitrification at shallow depths are responsible for early season N<sub>2</sub>O emissions under alternate wetting and drying management in an Italian rice paddy system. Soil Biology and

Biochemistry, 2018; 120: 58-69.

- [33] Zhang X, Zhang J Q, Song M X, Dong Y B, Xiong Z Q. N<sub>2</sub>O and NO production and functional microbes responding to biochar aging process in an intensified vegetable soil. Environmental Pollution, 2022; 307: 119491.
- [34] Abalos D, Sanchez-Martin L, Garcia-Torres L, van Groenigen J W, Vallejo A. Management of irrigation frequency and nitrogen fertilization to mitigate GHG and NO emissions from drip-fertigated crops. Science of the Total Environment, 2014; 490: 880–888.
- [35] Sánchez L, Arce A, Benito A, Garcia L, Vallejo A. Influence of drip and furrow irrigation systems on nitrogen oxide emissions from a horticultural crop. Soil Biology and Biochemistry, 2008; 40: 1698–1706.
- [36] Mehmood F, Wang G S, Gao Y, Liang Y P, Zain M, Rahman S, et al. Impacts of Irrigation Managements on Soil CO<sub>2</sub> Emission and Soil CH<sub>4</sub> Uptake of Winter Wheat Field in the North China Plain. Water, 2021: 13(15). doi:10.3390/W13152052.
- [37] Wang Y Q, Peng Z Y, Li Y C, Li Y L, Shu X T, Guo L P, et al. The effect of nitrification inhibitor on N<sub>2</sub>O, NO and N<sub>2</sub> emissions under different soil moisture levels in a permanent grassland soil. Soil Biology and Biochemistry, 2017; 113: 153–160.
- [38] Andrews H, Homyak P, Oikawa P, Wang J, Jenerette G. Water-conscious management strategies reduce per-yield irrigation and soil emissions of CO<sub>2</sub>, N<sub>2</sub>O, and NO in high-temperature forage cropping systems. Agriculture, Ecosystems & Environment, 2022; 332: 107944.
- [39] Wang G S, Liang Y P, Zhang Q, Shiva K, Gao Y, Shen X J, et al. Mitigated CH4 and N2O emissions and improved irrigation water use efficiency in winter wheat field with surface drip irrigation in the North China Plain. Agricultural Water Management, 2016; 163: 403–407.
- [40] Gu P, Jiao Y, Yang W Z, Wen H Y, Bai S G, Yang J. Effects of different irrigation methods on soil microbial abundance and permeability in farmland. Journal of Irrigation and Drainage, 2018; 37(1): 21–27.
- [41] Wei Y, Zhang C, Wang Y, Zhu J, Jin J. Effects of irrigation methods on soil water distribution and optimum irrigation schedule in apple orchard. Northern Horticulture, 2017; 11: 22–29.
- [42] Guo S F, Qi Y C, Luo X L, Liu C Y, Peng Q, Yan Z Q, et al. Effects of drip irrigation on soil CO<sub>2</sub> and N<sub>2</sub>O emissions and their global warming potentials of spring wheat field in arid region. Journal of Agro-Environment Science, 2016; 35(4): 792–800.
- [43] Lv H F, Zhou W W, Dong J, He S P, Chen F, Bi M H, et al. Irrigation amount dominates soil mineral nitrogen leaching in plastic shed vegetable production systems. Agriculture, Ecosystems and Environment, 2021; 317: 107474.
- [44] Su C X, Kang R H, Huang W T, Fang Y. Temporal patterns of N<sub>2</sub>O fluxes from a rainfed maize field in Northeast China. Frontiers in Environmental Science, 2021; 9: 668084.
- [45] Graham S L, Laubach J, Hunt J E, Mudge P L, Nunez J, Rogers G, et al. Irrigation and grazing management affect leaching losses and soil nitrogen balance of lucerne. Agricultural Water Management, 2022; 259: 107233.
- [46] Yang Y, Wang Z, Hu Y, Zeng Z, Feng H J, Yang Q G, et al. Irrigation frequency alters the abundance and community structure of ammoniaoxidizing archaea and bacteria in a northern Chinese upland soil. European Journal of Soil Biology, 2017; 83: 34–42.
- [47] Liao X, Niu Y H, Liu D Y, Chen Z M, He T H, Luo J F, et al. Four-year continuous residual effects of biochar application to a sandy loam soil on crop yield and N<sub>2</sub>O and NO emissions under maize-wheat rotation. Agriculture, Ecosystems and Environment, 2020; 302: 107109.
- [48] Omirou M, Stephanou C, Anastopoulos I, Philippot L, Ioannides I M. Differential response of N<sub>2</sub>O emissions, N<sub>2</sub>O-producing and N<sub>2</sub>O-reducing bacteria to varying tetracycline doses in fertilized soil. Environmental Research, 2022; 214: 114013.
- [49] Liu J Y, Qiu H S, Zhang W Z, Hu Z Z, Lv M C, Wang Y, et al. Mulched drip irrigation and biochar application reduce gaseous nitrogen emissions, but increase nitrogen uptake and peanut yield. Science of the Total Environment, 2022; 830: 154753.
- [50] Taylor A E, Myrold D D, Bottomley P J. Temperature affects the kinetics of nitrite oxidation and nitrification coupling in four agricultural soils. Soil Biology and Biochemistry, 2019; 136: 107523.
- [51] Hou M M, Shao X H, Jin Q, Gao X L. A N-15 tracing technique-based analysis of the fate of fertilizer N: a 4-year case study in eastern China. Arch Agron Soil Sci, 2017; 63(1): 74–83.